Original Article

The stock status of the pelagic fishes in Banyuasin coastal waters, Indonesia

Fauziyah¹ • Wike Ayu Eka Putri¹ • Anna Ida Sunaryo Purwiyanto¹ • Fitri Agustriani¹ • Ermatita² • Apriansvah Putra²

Correspondence

Dr Fauziyah; Marine Science Department, Faculty of Mathematics and Natural Sciences, Sriwijaya University, Ogan Ilir Regency, Province of South Sumatra, Indonesia



siti fauziyah@yahoo.com

Manuscript history

Received 10 November 2019 | Revised 18 May 2020 | Accepted 2 June 2020 | Published online 11 July 2020

Fauziyah, Putri WAE, Purwiyanto AlS, Agustriani F, Ermatita, Putra A (2020) The stock status of the pelagic fishes in Banyuasin coastal waters, Indonesia. Journal of Fisheries 8(2): 798-807.

Abstract

Information on the condition of pelagic fish stocks in Banyuasin coastal waters is currently very limited. This study aimed to estimate the stock status of Auxis spp., Scomberomorus commerson, Selaroides leptolepis and Rastrelliger spp. based on the time series catch and effort data from 2008 to 2016 in the Banyuasin coastal waters, South Sumatra Province, Indonesia. All seven surplus production models, model performance, and fish stocks status were estimated. In order to determine the best-fitted model, several indicators of model performance were required. The Pella and Tomlinson model was the best-fitted model for S. commerson while the best-fitted model for Rastrelliger spp., Auxis spp. and S. leptolepis was the Fox model. The optimum effort (E_{MSY}) value for S. commerson, Rastrelliger spp., Auxis spp. and S. leptolepis were 68677, 18226, 23402 and 22403 trips respectively. The maximum sustainable catch (C_{MSY}) value for S. commerson, Rastrelliger spp., Auxis spp. and S. leptolepis were 1845, 515, 286, 667 tons respectively. In 2016 the stock of S. commerson was in recovery condition whereas it was subjected to overfishing for Auxis spp. and depleted for S. leptolepis and Rastrelliger spp.

Keywords: Banyuasin; pelagic fish; stock status; surplus production model

1 | INTRODUCTION

Banyuasin coastal waters are the centre of capture fisheries in South Sumatra Province, Indonesia (Fauziyah et al. 2018a, 2019). Fishing units that develop in the area are included to the small-scale fisheries category (Fauziyah et al. 2018a). The local government has not applied regulatory methods to manage fisheries resources such as controlling fishing gear and technology, fishing time and area, as well as limiting fishing units. The regulations methods can be used to protect fisheries resources (Chae and Pascoe 2005).

Currently, fisheries statistics in Indonesia (including in Banyuasin) only record catch and effort data by each gear type, while data and information on effort level, exploita-

¹ Marine Science Department, Faculty of Mathematics and Natural Sciences, Sriwijaya University, Ogan Ilir Regency, Province of South Sumatra, Indonesia

² Information System Department, Faculty of Computer Science, Sriwijaya University, Ogan Ilir Regency, Province of South Sumatra, Indonesia

tion level, and fish stock status is not yet available. The data are very important for sustainable management of fisheries resources. Various researches are being carried out to reach the equilibrium between the populations of aquatic species and dynamically fluctuating and changing environments. Therefore, sustainable harvests are needed for determining how much fish stock can be sustainably taken from the fishery (Holmes et al. 2014). Two key factors that need to be balanced for sustainable fishing are the exploitation and the fishing effort levels (Fauziyah et al. 2020). Other influencing factors are predator abundance, food availability, environmental variables, climate change, and so on. The exploitation and fishing effort level can be estimated using a surplus production model (SPMs). When the data is limited, SPM can be used to estimate the maximum sustainable yield (MSY) and can assess fish stock (Chaloupka and Balazs 2007; Bordet and Rivest 2014).

The SPMs are the simplest stock-assessment models commonly used in fisheries (Walters and Hilborn 1976; Kurian 1989; Chen and Andrew 1998). These models only need a time-series data of catch and catch per unit of effort (CPUE) for running the models (Yoshimoto and Clarke 1993; Chen and Andrew 1998; Chen 2003) and relatively more available in most centres of fishing (Tinungki et al. 2004). These models can be used as an alternative analysis when virtual population analysis cannot be done due to the age structure information of the catch is not available (Meraz-Sánchez et al. 2013).

In order to better assess the dynamic fisheries resources, the approach and concept of SPM have developed by many authors; common SPMs are Schaefer's Model; Fox Model; Schnute Model; Gulland Model; Clark, Yoshimoto and Pooley (CYP) Model; Pella & Tomlinson Model; Walter-Hilborn Model; Cushing Model etc. (Tinungki *et al.* 2004; Kekenusa *et al.* 2018; Fauziyah *et al.* 2020). However, some researchers used several models to get the best-fitted models (Colvin *et al.* 2012; Mayalibit *et al.* 2014; Kekenusa *et al.* 2015, 2018; Singh 2015; Sin and Yew 2016). In the present analysis, seven different SPMs were applied to assess the current stock status of the common pelagic fishes including *Auxis* spp., *Scomberomorus commerson*, *Selaroides leptolepis* and *Rastrelliger* spp.

The status assessment of fish resource stocks in the Banyuasin coastal waters is poorly studied. The fish stock status in these waters has been only analysed for snapper in 2018 (Fauziyah *et al.* 2020) and the status of pelagic fish stocks remain unknown. The information on fish stock status are essential as basic data are important for determining the appropriate fisheries management and action plans. Therefore, the purpose of this study was to assess the status of pelagic fish stocks in the Banyuasin coastal waters using the best-fitted SPM based on the time series catch and effort data from 2008 to 2016.

2 | METHODOLOGY

2.1 Study area

This study was carried out in Banyuasin coastal waters (Figure 1) of the South Sumatra Province, Indonesia. This location has the most significant contribution to capture fisheries production in South Sumatra Province.

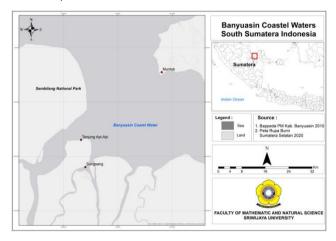


FIGURE 1 The study location of Banyuasin coastal water, Province of South Sumatra, Indonesia.

2.2 Data

As the simplest stock-assessment models, the SPMs only use annual fish catch and fishing effort data. In this study, all data were retrieved from the Annual Fishery Statistics of Banyuasin Regency (DKP 2009 – 2017). However, four species of pelagic fish (*Auxis* spp., *Scomberomorus commerson*, *Selaroides leptolepis* and *Rastrelliger* spp.) were considered in this study. Data of 2008 – 2016 were considered in this study and they were classified based on fish catch in fishing gears per fishing trip for each species. The fishing effort considered here is the number of operational fishing boats (*i.e.* trips) whereas the total catch represents the total amount of fish landed (Baset *et al.* 2017).

In Banyuasin coastal waters, the local fishermen captured *S. commerson* by using drift gillnet, set gillnet, trammel net, hook and lines, stationary lift net, and traps. *Auxis* spp. were captured using drift gillnet, trammel net, hook and lines, and traps. *Rastrelliger* spp. were captured using Danish seines, trammel net, stationary lift net, and traps. Whereas *S. leptolepis* were captured using Danish seines, drift gillnet, set gillnet, trammel net, and stationary lift net.

2.3 Catch per unit effort (CPUE) and effort standardization

As fishing gears used for sampling had different catchabilities they were standardised using the following equations (after Sparre and Venema 1998; King 2007; Fauziyah et al. 2018b):

$$E_{jt} = \varphi_{jt}D_{jt}$$

$$\varphi_{jt} = \frac{U_{jt}}{U_{st}}$$

$$U_{jt} = \frac{C_{jt}}{D_{jt}}$$

Where.

 E_{jt} = Effort from gear j at t standardised; D_{jt} = Effort from gear j at t period (trip); φ_{jt} = Fishing power of gear j at t period; U_{jt} = CPUE of gear j at t period; U_{st} = CPUE of gear based for standardised; U_{jt} = CPUE of gear j at t period (ton / trip); C_{jt} = the catch of gear j at t period (ton).

2.4 Surplus Production Models

In this study, the SPMs have used the catch data of each pelagic species and fishing effort used in term of the fishing trip number. The functions for seven SPMs equations are presented in Table 1. The sustainable catch of each pelagic species can be estimated by the logistic growth function and Gompertz growth function (Sin and Yew 2016). Parameters estimated from Schaefer, Gulland, Pella & Tomlinson, Walter and Hilborn, and Schnute models were used in the logistic catch equation while those estimated from Fox and Clarke Yoshimoto Pooley (CYP) models were used in Gompertz catch equation.

TABLE 1 The equations for Surplus Production Models and reference points.

| Model | Equation | MSY | References |
|------------------------|---|--|--|
| Schaefer | $\frac{c_t}{E_t} = \alpha - \beta E_t;$ $C_t = aE_t - bE_t^2$ | $E_{msy} = \frac{a}{2b}$ $C_{msy} = \frac{a^2}{4b}$ | Aristiantin <i>et al.</i> (2017); Kekenusa <i>et al.</i> (2018) |
| Gulland | $U_t = \frac{C_t}{\bar{E}_t} = a - b\bar{E}_t$ $C_t = a\bar{E}_t - b\bar{E}_t^2$ | $E_{msy} = \frac{a}{2b}$ $C_{msy} = \frac{a^2}{4b}$ | Ricker (1975); Widodo (1986); Singh (2015) |
| Pella & Tomlin- son | $U_t = \frac{C_t}{E_t} = a - bE_t^{m-1}$ $C_t = aE_t - bE_t^m$ | $E_{msy} = \left(\frac{a}{mb}\right)^{(1/(m-1)}$ $C_{msy} = aE_{msy} + bE_{msy}^{m}$ | Widodo (1986); Singh (2015) |
| Fox | $Ln\left(\frac{C_t}{E_t}\right) = a - bE_t$ $C_t = E_t Exp(a - bE_t)$ | $E_{msy} = \frac{1}{b}$ $C_{msy} = \frac{1}{b} exp(a-1)$ | Mohsin <i>et al.</i> (2017); Kekenusa <i>et al.</i> (2018) |
| Walters-Hilborn | $\begin{aligned} &\frac{U_{t+1}}{U_t} - 1 = a + bU_t + cE_t \\ &C_t = KqE_t - \frac{Kq^2}{r}E_t^2 \\ &\text{a=r; q=-c; K= a/(bc)} \end{aligned}$ | $E_{msy} = -\frac{a}{2c} = -\frac{r}{2q}$ $C_{msy} = \frac{a^2}{4b} = \frac{rK}{4}$ | Kekenusa <i>et al.</i> (2018) |
| Schnute | $Y_{t} = a + bX_{it} + cX_{2t}$ $C_{t} = KqE_{t} - \frac{Kq^{2}}{r}E_{t}^{2}$ $Y_{t}=Ln(U_{t+1}/U_{t}); X_{1t}=\frac{1}{2}(U_{t}+U_{t+1});$ $X_{2t}=\frac{1}{2}(E_{t}+E_{t+1});$ $a=r; q=-b; K=a/(bc)$ | $E_{msy} = -\frac{a}{2c} = -\frac{r}{2q}$ $C_{msy} = \frac{\alpha^2}{4b} = \frac{rK}{4}$ | Sholahuddin <i>et al.</i> (2015); Kekenusa <i>et al.</i> (2018) |
| СҮР | $\begin{split} Y_t &= a + b X_{it} + c X_{2t} \\ C_t &= Kq E_t exp \left(\frac{-q}{r} E_t \right) \\ Y_t &= \ln(U_{t+1}); X_{1t} = \ln(U_t); X_{2t} = (E_t + E_{t+1}); \\ a &= \hat{a} \ln(qK); r = 2(1-b)/(1+b) \\ q &= -c(2+r); K = e^Q/q \\ Q &= a(2+r)/(2r) \end{split}$ | $E_{msy} = \frac{r}{q}$ $C_{msy} = \frac{a^2}{4bc} = \frac{rK}{e}$ | Supriatna <i>et al.</i> (2016); Kekenusa <i>et al.</i> (2018) |

 E_t , effort standardised at t period; \bar{E}_t , moving average of effort standardise at t period; E_{t+1} , effort standardised at t+1 period; C_t , catch at t period; U_t , CPUE standardised at t period; U_t , intrinsic growth rate; q, catchability coefficient; K, carrying capacity; q, p and p are regression coefficients.

2.5 Best-fitted model

By using linear regression between CPUE and effort regression coefficient values (a, b and c) and biological parameters (r, q and K) can be obtained. The best-fitted of SPM is a model that has a sign suitability of biological parameter (positive value) and the best model performance among the other SPM applied.

For the equation of the Schaefer, Pella & Tomlinson, Fox, and Gulland models, only the intercept and slope values are obtained while the biological parameter can not be directly identified. For the Schaefer Model, Gulland Model, Pella & Tomlinson Models, the biological parameters are considered appropriate if the value of the intercepts (a) is positive and the slope (b) is negative. While for the Fox model, the value of slope (b) must be negative (Sparre and Venema 1998; Kekenusa et al. 2018). Only models that have the sign suitability can be carried out by the model performance test. Several researchers (e.g. Siyal et al. 2013; Seong et al. 2015; Singh 2015) used different models performance tests such as determination coefficient (R^2) , root means square error (RMSE), mean absolute deviation (MAD), mean square error (MSE), mean absolute percentage error (MAPE), RMSE-observations standard deviation ratio (RSR) and Nash-Sutcliffe efficiency (NSE). The best model is the model that has the highest R^2 and NSE values, and the contrary has the lowest MAD, MSE, RMSE, MAPE and RSR (Singh 2015; Fauziyah et al. 2020). These values must be standardised with the scoring method to obtain the same standard values so that it is easier to determine the best model. The standardization formulas (Iskandar and Guntur 2014; Fauziyah et al. 2018a) are as follows:

$$V(X) = \frac{X - X_0}{X_a - X_0}$$

$$V(A) = \sum_{i=a}^{n} Vi(Xi)$$

$$i = a, b, c, d, \dots, n$$

Where: V(X), value function of criteria X; X, value of criteria X; X_o , the best value of criteria X; X_o , the worst value of criteria X; V(A), value function of alternatives A; Vi(Xi), value function of alternatives in criteria i.

2.6 | Fish stock status

Different categories of fish stock status were considered by several researchers (see Carruthers et~al.~2012; Tsikliras et~al.~2015; Froese et~al.~2018 for details). The Indonesian government also makes an exploitation level classification that is referring to the estimation of potential, total allowable catches and exploitation level of fish resources in the Fisheries Management Areas within the country. Based on these, this study created a modification of the fish stock status (Table 2) with consideration of $C_{\rm MSY}$ and $E_{\rm MSY}$ as reference points (Fauziyah et~al.~2020).

TABLE 2 The classification of fish stock status.

| The status of fisherie | The fish stock | |
|---------------------------|----------------------|-------------------|
| Exploitation level | Fishing effort Level | status |
| Over-exploited | Underfishing | Healthy stock |
| $(C/C_{MSY} \ge 1)$ | $(E/E_{\rm MSY}<1)$ | |
| Over-exploited | Overfishing | Depleting stock |
| $(C/C_{MSY} \ge 1)$ | $(E/E_{MSY} \ge 1)$ | |
| Fully-exploited | Underfishing | Recovery stock |
| $(0.5 \le C/C_{MSY} < 1)$ | $(E/E_{\rm MSY}<1)$ | |
| Fully-exploited | Overfishing | Overfishing stock |
| $(0.5 \le C/C_{MSY} < 1)$ | $(E/E_{MSY} \ge 1)$ | |
| Moderate exploited | Overfishing | Overfishing stock |
| $(0.2 < C/C_{MSY} < 0.5)$ | $(E/E_{MSY} \ge 1)$ | |
| Moderate exploited | Underfishing | Transitional re- |
| $(C/C_{MSY} < 0.5)$ | $(E/E_{\rm MSY}<1)$ | covery stock |
| Moderate exploited | Overfishing | Collapsed stock |
| $(C/C_{MSY} \leq 0.2)$ | $(E/E_{MSY} \ge 1)$ | |

3 | RESULTS

3.1. Best fitted model

The best-fitted model for Rastrelliger spp., S. commerson, Auxis spp. and S. leptolepis using various SPM are presented in Tables 3–6 respectively. The results showed that Walter-Hilborn (Table 3) and Schnute (Tabel 4) model do not adequately fit for Rastrelliger spp. and S. commerson because the biological parameters did not show the proper sign. While Gulland, Walter-Hilborn, Schnute and CYP models were not adequately fit for Auxis spp (Table 5). Whereas Walter-Hilborn, Schnute and CYP models were not adequately fit for S. leptolepis (Table 6).

Table 3 shows that the Fox model was the best-fitted model for *Rastrelliger* spp. based on scoring value (V(A) = 7). The scoring method was carried out on seven parameters which indicated model performance, *viz.* R^2 , NSE, MAD, MSE, RMSE, MAPE, and RSR. The parameters values were 0.905, 0.24, 19.678, 630.831, 25.116, 0.039 and 0.872 respectively. The parameter values were the best value when compared to other types of SPM used in the analysis. The value of $E_{\rm MSY}$, $C_{\rm MSY}$ and Total Allowable Catch (TAC) were 18226 trips, 515 tons and 412 tons respectively.

For *S. commerson*, Pella and Tomlinson's model was the best-fitted model based on scoring value (V(A) = 6.972). The scoring values of seven model performance parameters were as follows: R^2 , 0.802; NSE, 0.771; MAD, 73.665; MSE 8077.282; RMSE 89.874; MAPE 0.050; and RSR 0.478. The Pella and Tomlinson model had the six best values of the seven model performance parameters (R^2 , NSE, MSE, RMSE, MAPE and RSR). While the best value for MAPE was owned by the CYP model. The values of $E_{\rm MSY}$, $C_{\rm MSY}$, and TAC for the Pella and Tomlinson model were 68677 trips, 1845 tons and 1476 tons respectively.

TABLE 3 Summary statistics for various SPMs for *Rastrelliger* spp. (SPM₁, Schaefer; SPM₂, Gulland; SPM₃, Fox; SPM₄, Pella and Tomlinson; SPM₅, Walter-Hilborn; SPM₆, Schnute; SPM₇, CYP; V(A), scoring value; NA, not appropriate).

| Parameter | Surplus Production Models (SPMs) | | | | | | | |
|------------------|----------------------------------|------------------|------------------|------------------|------------------------|------------------|------------------|--|
| | SPM ₁ | SPM ₂ | SPM ₃ | SPM ₄ | SPM₅ | SPM ₆ | SPM ₇ | |
| Sign suitability | | | | | | | | |
| а | 0.050760 | 0.053710 | -2.566806 | 0.300097 | -2.23425 | 1.02611 | -4.2251 | |
| b | -0.000001 | -0.000001 | -0.000055 | -0.101927 | 25.06502 | -19.36268 | -0.7079 | |
| С | | | | | 0.00008 | -0.00002 | -0.0001 | |
| r | | | | | -2.23425 ^{NA} | 1.02611 | 11.6942 | |
| K | | | | | -0.00008 NA | 0.00002 | 0.0007 | |
| q | | | | | -1129.783 NA | 2160.08689 | 122.0186 | |
| m | - | - | - | 1.4 | | | | |
| Performance to | est | | | | | | | |
| R^2 | 0.886 | 0.614 | 0.905 | 0.893 | | 0.009012 | 0.805 | |
| NSE | 0.149 | -0.025 | 0.240 | 0.225 | | -1.719757 | 0.202 | |
| MAD | 21.938 | 22.287 | 19.678 | 20.142 | | 39.25051845 | 19.807 | |
| MSE | 706.100 | 850.713 | 630.831 | 642.837 | | 2,257 | 661.984 | |
| RMSE | 26.573 | 29.167 | 25.116 | 25.354 | | 47.51 | 25.729 | |
| MAPE | 0.044 | 0.046 | 0.039 | 0.041 | | 0.0806 | 0.040 | |
| RSR | 0.922 | 1.013 | 0.872 | 0.880 | | 1.64917 | 0.893 | |
| MSY | | | | | | | | |
| E_{MSY} | 20324 | 19652 | 18226 | 18872 | | 20913 | 16935 | |
| C _{MSY} | 516 | 528 | 515 | 515 | | 554.12 | 525 | |
| TAC | 413 | 422 | 412 | 412 | | 443 | 420 | |
| Best-fitted mod | del | | | | | | | |
| V(A) | 6.521 | 6 | 7.000 | 6.900 | | 0 | 6.773 | |

TABLE 4 Summary statistics for various SPM of *Scomberomorus commerson* (SPM₁, Schaefer; SPM₂, Gulland; SPM₃, Fox; SPM₄, Pella and Tomlinson; SPM₅, Walter-Hilborn; SPM₆, Schnute; SPM₇, CYP; *V(A)*, scoring value; NA, not appropriate).

| Parameter | Surplus Production Models (SPMs) | | | | | | | |
|------------------|----------------------------------|------------------|------------------|------------------|------------------|------------------------|------------------|--|
| | SPM ₁ | SPM ₂ | SPM ₃ | SPM ₄ | SPM ₅ | SPM ₆ | SPM ₇ | |
| Sign suitability | | | | | | | | |
| а | 0.070833 | 0.068935 | -2.558363 | 0.295441 | 1.86391 | -2.69039 | -4.9302 | |
| b | -0.000001 | -0.000001 | -0.000016 | -0.088185 | -29.73651 | 36.02756 | -0.9252 | |
| С | | | | | -0.00002 | 0.00003 | -0.0000 | |
| r | | | | | 1.86391 | -2.69039 ^{NA} | 51.4714 | |
| K | | | | | 0.00002 | -0.00003 NA | 0.0008 | |
| q | | | | | 4005.81323 | –2536.56 ^{NA} | 95.5593 | |
| m | - | - | - | 1.1 | | | | |
| Performance test | | | | | | | | |
| R^2 | 0.776 | 0.513 | 0.798 | 0.802 | 0.442446 | | 0.796 | |
| NSE | 0.753 | 0.755 | 0.766 | 0.771 | 0.716233 | | 0.766 | |
| MAD | 74.028 | 73.595 | 73.665 | 73.591 | 80.483112 | | 73.395 | |
| MSE | 8716.116 | 8632.994 | 8271.787 | 8077.282 | 10014 | | 8262.571 | |
| RMSE | 93.360 | 92.914 | 90.949 | 89.874 | 100.07 | | 90.899 | |
| MAPE | 0.05085 | 0.05011 | 0.05013 | 0.05000 | 0.05370 | | 0.05006 | |
| RSR | 0.4970 | 0.4946 | 0.4842 | 0.4784 | 0.5327 | | 0.4839 | |
| MSY | | | | | | | | |
| E_{MSY} | 48108 | 49658 | 63173 | 68677 | 59559 | | 63681 | |
| C_{MSY} | 1704 | 1712 | 1800 | 1845 | 1867 | | 1809 | |
| TAC | 1363 | 1369 | 1440 | 1476 | 1493 | | 1448 | |
| Best model | | | | | | | | |
| V(A) | 5.266 | 5 | 6.505 | 6.972 | 0 | | 6.577 | |

TABLE 5 Summary statistics for various SPMs of *Auxis* spp. (SPM₁, Schaefer; SPM₂, Gulland; SPM₃, Fox; SPM₄, Pella and Tomlinson; SPM₅, Walter-Hilborn; SPM₆, Schnute; SPM₇, CYP; *V(A)*, scoring value; NA, not appropriate).

| Parameter | Surplus Production Models (SPMs) | | | | | | | |
|------------------|----------------------------------|-----------------------|------------------|------------------|------------------------|------------------------|------------------|--|
| | SPM ₁ | SPM ₂ | SPM ₃ | SPM ₄ | SPM ₅ | SPM ₆ | SPM ₇ | |
| Sign suitability | | | | | | | | |
| а | 0.024629 | 0.008586 | -3.405611 | 0.152247 | -2.03121 | -3.38829 | -3.9218 | |
| b | -0.000001 | 0.00000 ^{NA} | -0.000043 | -0.051071 | -14.84577 | 45.81594 | 0.3805 | |
| С | | | | | 0.00008 | 0.00011 | 0.0000 | |
| r | | | | | -2.03121 ^{NA} | -3.38829 NA | 0.8975 | |
| K | | | | | -0.00008 NA | -0.00011 NA | -0.0001 NA | |
| q | | | | | 1693.058 NA | –691.586 ^{NA} | -30.604 NA | |
| m | - | - | - | 1.2 | | | | |
| Performance to | est | | | | | | | |
| R^2 | 0.318 | - | 0.267 | 0.341 | - | - | - | |
| NSE | 0.869 | - | 0.875 | 0.873 | - | - | - | |
| MAD | 35.789 | - | 34.963 | 35.215 | - | - | - | |
| MSE | 1335.751 | - | 1278.131 | 1297.837 | - | - | - | |
| RMSE | 36.548 | - | 35.751 | 36.026 | - | - | - | |
| MAPE | 0.129 | - | 0.125 | 0.127 | - | - | - | |
| RSR | 0.362 | - | 0.354 | 0.357 | - | - | - | |
| MSY | | | | | | | | |
| E_{MSY} | 23,693 | - | 23,402 | 1,916 | - | - | - | |
| C_{MSY} | 292 | - | 286 | 83 | - | - | - | |
| TAC | 233 | - | 229 | 67 | - | - | - | |
| Best model | Best model | | | | | | | |
| V(A) | 0.688 | - | 6.000 | 4.825 | - | - | - | |

Table 6 Summary statistics for various SPMs of *Selaroides leptolepis* (SPM₁, Schaefer; SPM₂, Gulland; SPM₃, Fox; SPM₄, Pella and Tomlinson; SPM₅, Walter-Hilborn; SPM₆, Schnute; SPM₇, CYP; *V(A)*, scoring value; NA, not appropriate).

| Parameter | Surplus Production Models (SPMs) | | | | | | |
|------------------|----------------------------------|------------------|------------------|------------------|------------------------|------------------|-----------------------|
| | SPM ₁ | SPM ₂ | SPM ₃ | SPM ₄ | SPM₅ | SPM ₆ | SPM ₇ |
| Sign suitability | • | | | | | | |
| а | 0.068603 | 0.031874 | -2.51469 | 0.401944 | -0.47293 | -0.58155 | -2.9094 |
| b | -0.000002 | -0.000003 | -0.000045 | -0.136186 | -6.46752 | -12.50757 | 0.2501 |
| С | | | | | 0.00002 | 0.00003 | 0.0000 |
| r | | | | | -0.47293 ^{NA} | -0.58155 NA | 1.1996 |
| K | | | | | -0.00002^{NA} | -0.00003 NA | -0.00001^{NA} |
| q | | | | | 2975.958 | 1433.1421 | -3330.1 ^{NA} |
| m | - | - | - | 1.1 | | | |
| Performance t | est | | | | | | |
| R^2 | 0.583 | 0.179 | 0.650 | 0.641 | - | - | - |
| NSE | -0.617 | -1.346 | -0.212 | -0.338 | - | - | - |
| MAD | 110.898 | 94.414 | 89.082 | 99.716 | - | - | - |
| MSE | 20716.965 | 30046.535 | 15520.150 | 17137.178 | - | - | - |
| RMSE | 143.934 | 173.339 | 124.580 | 130.909 | - | - | - |
| MAPE | 0.182 | 0.129 | 0.142 | 0.166 | - | - | - |
| RSR | 1.272 | 1.532 | 1.101 | 1.157 | - | - | - |
| MSY | | | | | | | |
| E_{MSY} | 21172 | 54783 | 22403 | 19337 | - | - | - |
| C_{MSY} | 726 | 873 | 667 | 707 | - | - | - |
| TAC | 581 | 698 | 533 | 565 | - | - | - |
| Best model | | | | | | | |
| V(A) | 3.349 | 1.76 | 6.763 | 5.323 | - | - | - |

In Table 5, the Fox model was the best-fitted model with the highest scoring value (6) for *Auxis* spp. Whereas for *S. leptolepis*, it was also the Fox model, obtained as the best-fitted model with the highest scoring value (6.763) (Table 6).

3.2 Fish stock status

Based on available data, the stock status of *S. leptolepis*, *Auxis* spp., *S. commerson* and *Rastrelliger* spp. in 2016 were depleting, overfishing, recovery and depleting stock respectively (Figure 2). The fisheries development of *Rastrelliger* spp. between 2008 and 2016 fluctuated (Figure 3). *Rastrelliger* catch in the 2008 – 2012 period did not exceed the sustainable catches ($C_{\rm MSY}$) but it was over the optimum efforts ($E_{\rm MSY}$) that results in an overfishing condition of the stock. Furthermore, in 2013 there was a decreasing effort below $E_{\rm MSY}$ value, but the catch was just higher than $C_{\rm MSY}$ value that indicates a healthy condition of the stock. In 2014 – 2016, increased fishing efforts beyond $E_{\rm MSY}$ value also results in more catches than $C_{\rm MSY}$ value put the stock in depleting condition.

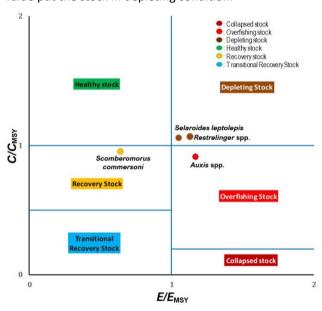


FIGURE 2 The stock status of four pelagic fishes in Banyuasin coastal waters for the 2016 year based on data from 2008 – 2016).

The fisheries development of *S. commerson* as shown in Figure 4 which indicates an increase in the catch in 2008 – 2016 period but it did not exceed $C_{\rm MSY}$ value (fully exploited) and $E_{\rm MSY}$ value (*i.e.* underfished). These conditions indicated that the status of fish stocks was in recovery condition. The catch of *Auxis* spp. exceeded both the $C_{\rm MSY}$ value (over-exploited) and $E_{\rm MSY}$ value (*i.e.* overfished) in 2011 – 2013 period that put the stock status in a depleting condition (Figure 5). Furthermore, in the 2014 – 2016, the efforts carried out exceeded the $E_{\rm MSY}$ value (*i.e.* overfished) but the catch was below $C_{\rm MSY}$ (*i.e.* fully exploited) that results in an overfishing condition.

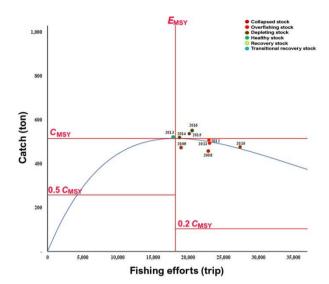


FIGURE 3 Fitted equilibrium Fox model and fish stock status for *Rastrelliger* spp. in Banyuasin coastal waters.

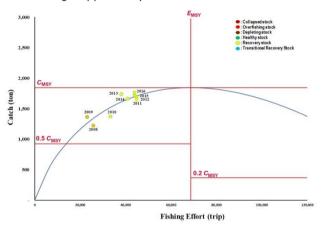


FIGURE 4 Fitted equilibrium Pella & Tomlinson model and fish stock status for *Scomberomorus commerson* in Banyuasin coastal waters.

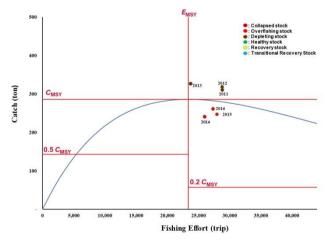


FIGURE 5 Fitted equilibrium Fox model and fish stock status for *Auxis* spp. in Banyuasin coastal waters.

The catch of *S. leptolepis* in 2008 exceeded the $C_{\rm MSY}$ (*i.e.* overexploited) but the effort did not exceed the $E_{\rm MSY}$ (*i.e.* underfished) reflecting a healthy condition of the stock

(Figure 6). In 2009 increased efforts were recorded but they did not exceed the $E_{\rm MSY}$ (i.e underfished) but the catch decreased dramatically below the $C_{\rm MSY}$ value (i.e. fully exploited). This phenomenon indicated that the stock status for the S. leptolepis in 2009 was recovering phase (recovery stock). Furthermore, in 2010 – 2013, the trend of efforts have increased and exceeded $E_{\rm MSY}$ value but the catch obtained did not exceed $C_{\rm MSY}$ value indicating an overfishing status of the stock. In 2014 – 2016, the efforts and catch exceeded both $E_{\rm MSY}$ and $C_{\rm MSY}$ values which indicated that the fish stock status was in depleting state.

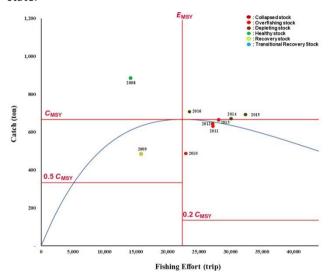


FIGURE 6 Fitted equilibrium Fox model and fish stock status for *Selaroides leptolepis* in Banyuasin coastal waters.

4 | DISCUSSION

This study has analysed seven SPMs and determined the best-fitted model based on R², NSE, MAD, MSE, RMSE, MAPE and RSR values using a scoring approach. Fox model was the best-fitted model for Rastrelliger spp. but the performance rating of the model was unsatisfactory (NSE = $0.24 \le 0.50$ and RSR = 0.872 > 0.70; Moriasi et al. 2007). Nonetheless, based on R^2 value (0.905 > 0.85) this indicated excellent performances (Makungo and Odiyo 2017). Fox's model was also the best-fitted model for Auxis spp. and S. leptolepis. Referring to NSE and RSR values, the model performance rating for Auxis spp. was very good $(0.75 \le NSE = 0.875 \le 1$; and $0 \le RSR = 0.354 \le 0.5$; (Moriasi et al. 2007) but R^2 value (0.267 < 0.5) indicated unsatisfactory performances (Makungo and Odiyo 2017). However, the model performance rating for S. leptolepis was unsatisfactory (NSE = $-0.212 \le 0.5$; RSR = 1.157 > 0.7; and $R^2 = 0.267 < 0.5$; Moriasi et al. 2007, Makungo and Odiyo 2017). On the contrary, the best-fitted model for S. commerson was Pella and Tomlinson model where the model performance rating was very good (0.75 ≤ NSE = $0.771 \le 1$); RSR (0 \le RSR = 0.478 \le 0.5; and 0.75 \le R² = 0.802 ≤ 0.85) (Moriasi et al. 2007; Makungo and Odiyo

2017). The results showed that the performance rating, based on RSR and NSE, yielded similar result but different for R^2 values. In this study, R^2 described the degree of collinearity between CPUE and effort (linear regression model) and did not describe the degree of collinearity between the catch and effort (equilibrium model) while the other parameters (NSE, MAD, MSE, RMSE, MAPE and RSR) described the model performance evaluation in terms of the equilibrium model. However, there is no firm consensus on acceptable model performance parameters and no single statistic can be used to assess all aspects of model performance (Duda *et al.* 2012; Seong *et al.* 2015).

The stock status of S. leptolepis, Auxis spp., S. commerson and Rastrelliger spp. in 2016 were in depleting, overfishing, recovery, and depleting conditions respectively. This condition indicated that fishing efforts for S. leptolepis, Auxis spp. and Rastrelliger spp. were larger than their estimated E_{MSY} . However, overfishing stock status was recorded for Auxis spp. which is in line with the assessment of same species in Talaud waters (Kekenusa et al. 2015) and Bitung waters of North Sulawesi (Kekenusa et al. 2018) of Indonesia. While depleting stock for S. leptolepis and Rastrelliger spp. in this study were different from the assessment conducted in the Karangantu National fishing port of Banten and Sunda Strait waters, where the results showed that an overfishing stock of the species (Mayalibit et al. 2014; Sarasati et al. 2016). Recovery stock for S. commerson in this study was different from the assessment in the Meranti Islands waters where the results have shown a depleting status (Syaputra et al. 2016). Thus, fishing effort and catch should be kept limited in order to obtain a maximum sustainable yield. Variation in catch depends not only on efforts but also on the environmental factors (Meraz-Sánchez et al. 2013). Thus, in addition to promoting the development of sustainable fishing grounds lowering the number of fishing vessels may yield an improvement of the overfished stocks (Chae and Pascoe 2005; Siyal et al. 2013). And it is important to monitor the fish stock status on regular basis (Meraz-Sánchez et al. 2013).

ACKNOWLEDGEMENT

We would like to thank the Ministry of Research, Technology, and Higher Education of the Republic of Indonesia for the support of PSN Institusi-Kemenristekdikti funds (Grant No. SP DIPA-042.06.1.401516/2019) in 2019. We are also grateful to the anonymous reviewers of the final version of the manuscript. Special thanks to Sriwijaya University and Department of Fisheries and Marine, Banyuasin Regency, Province of South Sumatera, Indonesia for their help.

CONFLICT OF INTEREST

The author declares no conflict of interest.

DATA AVAILABILITY STATEMENT

The data supporting the findings of this study are available within the article [and/or] its supplementary materials.

REFERENCES

- Aristiantin Y, Purwiyanto AIS, Fauziyah (2017) Stock assessment of skipjack tuna (*Katsuwonus pelamis*) at Nizam Zachman Ocean fishing port using FAO-ICLARM stock assessment tools. Jurnal Maspari 9: 43–50 [in Indonesian].
- Baset A, Qun L, Pavase TR, Hameed A, Niaz Z (2017) Estimation of maximum sustainable yield of *Scomberomorus* species fish stocks in Pakistan using surplus production models. Indian Journal of Geo-Marine Sciences 46: 2372–2378.
- Bordet C, Rivest LP (2014) A stochastic Pella Tomlinson model and its maximum sustainable yield. Journal of Theoretical Biology 360: 46–53.
- Carruthers TR, Walters CJ, McAllister MK (2012) Evaluating methods that classify fisheries stock status using only fisheries catch data. Fisheries Research 119–120: 66–79.
- Chae D, Pascoe S (2005) Use of simple bioeconomic models to estimate optimal effort. Aquatic Living Resources 18: 93–101.
- Chaloupka M, Balazs G (2007) Using Bayesian state-space modelling to assess the recovery and harvest potential of the Hawaiian green sea turtle stock. Ecological Modelling 205: 93–109.
- Chen Y (2003) Quality of fisheries data and uncertainty in stock assessment. Scientia Marina 67: 75–87.
- Chen Y, Andrew N (1998) Parameter estimation in modelling the dynamics of fish stock biomass: are currently used observation-error estimators reliable? Canadian Journal of Fisheries and Aquatic Sciences 55: 749–760.
- Colvin ME, Pierce CL, Stewart TW (2012) Semidiscrete biomass dynamic modeling: an improved approach for assessing fish stock responses to pulsed harvest events. Canadian Journal of Fisheries and Aquatic Sciences 69: 1710–1721.
- DKP (2009 2017) Buku statistik perikanan tahun 2008–2016. Dinas Kelautan dan Perikanan Provinsi Sumatera Selatan, Palembang.
- Duda PB, Hummel PR, Imhoff JC (2012) Basins/hspf: model use, calibration, and validation. Transactions of the ASABE 55: 1523–1547.
- Fauziyah, Agustriani F, Melda Situmorang D, Suteja Y (2018b) Fishing seasons of fish landed at Sungailiat archipelago fishing port in Bangka Regency. E3S Web of Conferences 47: 1–10.
- Fauziyah, Agustriani F, Purwiyanto AlS, Putri WAE, Suteja Y (2019) Influence of environmental parameters on the shrimp catch in Banyuasin. Journal of Physics: Confer-

- ence Series 1282 01210.
- Fauziyah, Agustriani F, Satria B, Putra A, Nailis W (2018a) Assessment of multigear type at small-scale fisheries in Sungsang Estuary Banyuasin District South Sumatra. Marine Fisheries 9: 183–197 (in Indonesian).
- Fauziyah, Purwiyanto AIS, Agustriani F, Putri WAE, Ermatita, Putra A (2020) Determining the stock status of snapper (*Lutjanus* sp.) using surplus production model: a case study in Banyuasin coastal waters, South Sumatra, Indonesia. IOP Conference Series: Earth and Environmental Science 404 012009.
- Froese R, Winker H, Coro G, Demirel N, Tsikliras AC, ... Matz-Lück N (2018) Status and rebuilding of European fisheries. Marine Policy 93: 159–170.
- Holmes L, Strauss K, Vos K de, Bonzon K (2014) Towards investment in sustainable fisheries: a framework for financing the transition. In: Environmental Defense Fund and The Prince of Wales's International Sustainability Unit. 82 pp.
- Iskandar D, Guntur A (2014) Technical and economic efficiency of Garuk fishing gear and their development opportunities in Rawameneng village, Subang Regency. Maspari Journal 6: 81–97 (in Indonesian).
- Kekenusa JS, Paendong MS, Weku WCD, Rondonuwu SB (2015) Determination of the status of utilization and management scenarios bonito (*Auxis rochei*) caught in the Talaud waters North Sulawesi. Science Journal of Applied Mathematics and Statistics 3: 39–46.
- Kekenusa SJ, Rondonuwu SB, Paendong MS (2018) Determination of the status of utilization and management scenarios bonito (*Auxis rochei*) caught in the Talaud Waters North Sulawesi. International Journal of ChemTech Research 3: 39.
- King M (2007) Fisheries biology, assessment and management, 2nd edition. Blackwell Publishing Ltd, Oxford.
- Kurian A (1989) Application of synthetic models for the assessment of bombay duck, *Harpodon Nehereus* (Ham.). Indian Journal of Fisheries 36: 275–283.
- Makungo R, Odiyo JO (2017) Estimating groundwater levels using system identification models in Nzhelele and Luvuvhu areas, Limpopo Province, South Africa. Physics and Chemistry of the Earth 100: 44–50.
- Mayalibit DNK, Kurnia R, Yonvitner (2014) Bioeconomic analysis for management of yellowstripe scad (*Selaroides leptolepis*, Cuvier and Valenciennes) landed in Karangantu, Banten. Bonorowo Wetlands 4: 49–57 (in Indonesian).
- Meraz-Sánchez R, Madrid-Vera J, Cisneros-Mata MÁ, Herrera DC (2013) An approach to assessment to population of the brown shrimp, *Farfantepenaeus californiensis* (Holmes, 1900), as a management fisheries tool in the Southeastern Gulf of California. Open Journal of Marine Science 03: 40–47.
- Mohsin M, Mu YT, Sun Z, Afsheen S, Memon AM, Kalhoro MT, Shah SBH (2017) Application of non-equilibrium

- models to evaluate fishery status of squids in Pakistani Marine Waters. Journal of Animal and Plant Sciences 27: 1031–1038.
- Moriasi DN, Arnold JG, Liew MW Van, Bingner RL, Harmel RD, Veith TL (2007) Model evaluation guidelines for systematic quantification of accuracy in watershed simulations. Transactions of the ASABE 50: 885–900.
- Ricker WE (1975) Computation and interpretation of biological statistics of fish populations. Bulletin of the Fisheries Research Board of Canada 191: 1–382.
- Sarasati W, Boer M, Sulistiono (2016) Status of stock *Rastrelliger* spp. as the base for fisheries management. Jurnal Perikanan Universitas Gajah Mada 18: 73–81 (in Indonesian).
- Seong C, Herand Y, Benham BL (2015) Automatic calibration tool for hydrologic simulation program-FORTRAN using a shuffled complex evolution algorithm. Water 7: 503–527.
- Sholahuddin A, Ramadhan AP, Supriatna AK (2015) The application of ANN-Linear Perceptron in the development of DSS for a fishery industry. Procedia Computer Science 72: 67–77.
- Sin MS, Yew TS (2016) Assessing the exploitation status of marine fisheries resources for the west coast of peninsular Malaysia trawl fishery. World Journal of Fish and Marine Sciences 8: 98–107.
- Singh NO (2015) Surplus production models with auto correlated errors. International Journal of Fisheries and Aquatic Studies 2: 217–220.
- Siyal FK, Li Y, Gao T, Liu Q (2013) Maximum sustainable yield estimates of silver pomfret, *Pampus argenteus* (Family: Strometidae) fishery in Pakistan. Pakistan Journal of Zoology 45: 447–452.
- Sparre P, Venema SC (1998) Introduction to tropical fish stock assessment. Part 1. Manual, FAO Fisheries Technical Paper 306.1, Rev. 2. FAO, Rome. 407 pp.
- Supriatna AK, Sholahuddin A, Ramadhan AP and Husniah H (2016) SOFish ver. 1.2 a decision support system for fishery managers in managing complex fish stocks. IOP Conference Series: Earth and Environmental Science 31: 012005.
- Syaputra E, Rifardi, Nasution S (2016) Preliminary study of bioecology of *Hemibgrus wyckii*, Bagridae. Jurnal Perikanan dan Kelautan 21: 1–8 (in Indonesian).
- Tinungki GM, Boer M, Monintja DR, Widodo J (2004) The Surshing Model: an hybrid model between surplus production model and cushing model for fish stock assessment estimation (case study: Lemuru fisheries in Bali strait)]. Jurnal Ilmu-ilmu Perairan dan Perikanan Indonesia 11: 135–138 (in Indonesian).
- Tsikliras AC, Dinouli A, Tsiros VZ, Tsalkou E (2015) The Mediterranean and Black Sea fisheries at risk from overexploitation. PLOS ONE 10: e0121188.
- Walters CJ, Hilborn R (1976) Adaptive control of fishing systems. Journal of the Fisheries Research Board of Cana-

da 33: 145-159.

- Widodo J (1986) Fox model and generalized production model another versions of surplus production models. Oseana 11: 143–149.
- Yoshimoto SS, Clarke RP (1993) Comparing dynamic versions of the Schaefer and Fox production models and their application to lobster fisheries. Canadian Journal of Fisheries and Aquatic Sciences 50: 181–189.

CONTRIBUTION OF THE AUTHORS

FA data collection;

AP data analysis;

F data analysis and manuscript preparation;

AISP manuscript preparation;

WAEP data collection and research supervision;

E research supervision.



Fauziyah https://orcid.org/0000-0003-3624-7956

WAE Putri https://orcid.org/0000-0002-1456-3088

AIS Purwiyanto https://orcid.org/0000-0002-9148-1713

F Agustriani https://orcid.org/0000-0003-2198-7687

Ermatita https://orcid.org/0000-0002-5267-0589

A Putra https://orcid.org/0000-0003-3062-7202